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USING FRESHWATER WETLANDS FOR WASTEWATER MANAGEMENT IN NORTH AMERICA

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ABSTRACT

The potential for using freshwater wetlands for wastewater management has received much attention in North America during recent years. Experimental studies have been conducted in several types of wetlands, mostly in the central and eastern United States. Several types of natural and artificial wetlands can be used for short-term treatment of wastewater, but the long-term effects are unknown for most types of wetlands. Only peat-based systems appear to be capable of long-term efficient processing of domestic wastewater, although the use of artificial pond-wetland systems appears to have the most potential and should be considered as an alternative to using natural wetlands. Other types of freshwater wetlands have limited ability to remove nutrients from wastewater. Because there is much variation between wetlands that have been studied, it appears that each situation is unique and must be considered individually with pilot studies being conducted before any wetland is used for wastewater processing.

INTRODUCTION

Since publication of the proceedings from a conference on biological control of water pollution (Tourbier and Pierson 1976), there have been several meetings during which the topic of using freshwater wetlands for treating wastewater, primarily domestic sewage, has been considered (Tilton et al. 1976, Am. Soc. Civil Eng. 1978). As a result, there are several review papers (Sloey et al. 1978, Kadlec 1980, Whigham and Bayley 1980, van der Valk et al. 1980) as well as several papers that have reported results of experiments on the ability of specific artificial and natural freshwater wetlands to assimilate wastewater (Tilton and Kadlec 1979, Whigham et al. 1980, Ewel and Odum 1978, Fetter et al. 1978, Boyt et al. 1977). Large-scale experimental wastewater projects have been conducted in Florida (Odum and Ewel 1978) and Michigan (Kadlec 1975, Tilton et al. 1976). While other projects, more modest in scale, have been conducted in Florida (Zoltek et-al. 1979), New Jersey (Whigham et al. 1980), Minnesota (Stanlick 1976), Wisconsin (Spangler et al. 1976), and New York (Small 1976, Woodwell 1977, Woodwell et al. 1974). Several other freshwater wetlands have been used to treat wastewater even though they were not initially designed as part of experimental studies (Kadlec 1980, Boyt et al. 1977, Grant and Patrick 1970). One question that needs to be asked at this point is whether or not any general conclusions have been reached from these studies. The purpose of this paper is not to provide a review of each of the sudies that have been conducted but to present a framework by which past experiences can be used to evaluate some general questions about whether or not natural or artificial wetlands can be used as design components in systems that are used for secondary and tertiary treatment of wastewater.

PREVIOUS STUDIES

Studies of wastewater additions to natural peat-based wetlands in Michigan (Tilton and Kadlec 1979, Kadlec 1980) and Florida (Zoltek et al. 1979) as well as Cypress wet-

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lands in Florida (Odum and Ewel 1978, Nessel 1978) have shown that those types of of ecosystems can efficiently process wastewater. Artificial wetlands in Wisconsin (Spangler et al. 1976), peat-based systems in Minnesota (Stanlick 1976), combined marsh-pond systems in California (Jokela and Jokela 1978) and New York (Small 1976, Woodwell 1977, Wolverton et al. 1976), and wetlands developed on dredge spoil material (Lee et al. 1975 and 1976) have been studied. All of these systems, with the exception of the gravel-based wetlands, appear to be able to process wastewater and the marsh-pond systems constructed in New York were shown to be particularly efficient at removing nutrients from wastewater.

Many of the artificial systems were based on earlier work of Seidel and coworkers in Europe (Czerwanda and Seidel 1965, Seidel 1976) which has resulted in the use of man-made wetlands to treat wastes in Holland, Hungary, Poland, and Yugoslavia. A commercial version of the systems used by Seidel was patented in the United States (Wagner 1974).

FACTORS DETERMINING USE OF WETLANDS

The most important primary factors that determine whether wetlands can be used as part of large-scale operations (e.g., more than a million gallons of wastewater per day), as part of small-scale operations (less than 1 million gallons per day), or not used at all, are:

- a. hydrologic characteristics, especially the turnover rate of water (i.e., the contact time)
- b. type of substrate and ability of macrophytes to utilize nutrients applied in wastewater, and
- c. seasonality.

Hydrologic characteristics are poorly understood for most types of wetlands (Greeson et al. 1980) yet are undoubtedly very important. In situations where water moves quickly through the wetland, one can expect minimal uptake of nutrients (Tilton and Kadlec 1979, Whigham et al. 1980) compared to situations where water would have a long residence time (Zoltek et al. 1979, Ewel and Odum 1978). As has been noted by van der Valk et al. (1980) and Tilton and Kadlec (1979), there are well-defined channels in many types of wetlands and the water is carried through the wetland with very little contact with the wetland surface. This situation would be particularly common in palustrine systems. On the other hand, peat-based systems seem to be quite appropriate because water usually has a long residence time in the substrate which increases the chances for sorption or assimilation (Tilton and Kadlec 1979, Zoltek et al. 1979). Even if the emergent plants did not efficiently remove nutrients, their presence would cause a decline in water velocity and would enhance the possibilities for sediment flocculation, and sorption in the litter and substrate (Boto and Patrick 1980). In all instances where efficient nutrient removal has occurred, influent water had a long residence time in the system.

One advantage of man-made wetlands would be the potential to control hydrologic characteristics. For example, in temperate latitudes where there is a strong pulse of nutrient uptake during the growing season and release usually associated with decomposition in the fall (Simpson et al. 1978) or losses during the spring thaw (Sloey et al. 1978), water could be processed during the assimilation period and the wetland could be treated as an open system while during the release period water could be retained or released at a much slower rate (Sloey et al. 1978).

Substrate conditions appear to be important in determining the capability of specific wetlands to efficiently treat wastewater (Whigham and Bayley 1980) although much more information is needed. Almost all types of wetlands appear to be able to process nutrients for some part of the year (van der Valk et al. 1980). For nitrogen, denitrification appears to be very important and may account for the notion that many wetlands may have an infinite capacity to process nitrogen (Patrick and Reddy 1976). For phosphorus, the patterns are not as clear. Tilton and Kadlec (1979), Zoltek et al. (1979), and Ewel and Odum (1978) found efficient phosphorus removal in the peat-based systems that they studied. Some of the phosphorus was accounted for by increased phosphorus uptake in the vegetation (Ewel and Odum 1978), with the remainder having been sorbed or otherwise immobilized in the substrate. It has been analytically very difficult to account for all phosphorus that was added and several authors did not find significant increases in substrate phosphorus. In nonpeat-based wetlands (Spangler et al. 1976, Fetter et al. 1978, Simpson et al. 1978), phosphorus may be retained seasonally, but there appears to be little long-term substrate storage and most display strong seasonal release patterns.

In nonpeat-based wetlands, immobilization appears to occur primarily in the litter zone (Whigham et al. 1980), although several authors (Klopatek 1978, Prentki et al. 1978, Kitchens et al. 1975) have suggested that macrophytes are very important. Consequently, it has been suggested that harvesting of macrophytes would increase the nutrient assimilation capacity of wetlands. It appears, however, that only a small amount of N and P could be removed by harvesting emergent (Spangler et al. 1976, Zoltek et al. 1979) or submerged (McNabb 1976) vegetation. The best possibilities for using wetland plants for nutrient removal appear to occur in situations where the nutrients are stored in woody plants (Ewel and Odum 1978).

Seasonal patterns of nutrient immobilization in wetlands must be considered. Wetlands in subtropical areas may be capable of processing wastewater throughout the year (Zoltek et al. 1979, Boyt et al. 1977, Odum and Ewel 1978). In areas where the wetland surface freezes, it is only possible to use wetlands during the summer months (Spangler et al. 1976). It should be noted, however, that eutrophication problems are not as bad during the winter months, and Kadlec (1980) listed one Michigan site as being operational for the entire year even though winters are severe in that part of the country.

DESIRABILITY OF TREATMENT BY WETLANDS

Other scientists have recommended against the widespread use of freshwater wetlands for wastewater management. Some wetlands may be capable of wastewater polishing but cannot be used for cost-effective treatment of large quantities of wastewater. These recommendations have been based on results of experimentation with wastewater addition to freshwater tidal wetlands in New Jersey (Whigham et al. 1980) and gravel-based artificial systems in Wisconsin (Spangler et al. 1978). Studies of nutrient dynamics in various types of wetlands (e.g., Correll et al. 1975), or a concern about the expanded use of wetlands for those purposes when much more information is needed about functional processes in wetlands has also led some authors to suggest that wetlands should not be used for wastewater management (van der Valk et al. 1980, Greeson et al. 1980).

LONG-TERM CONSEQUENCES OF WASTEWATER ADDITIONS

Most of the research projects that have been cited, were conducted for relatively short periods of time. We still do not know how long particular wetlands can be used to process wastewater before their efficiency declines. Table 1 lists data reported in

TABLE 1. Typical results of wetland tertiary treatment in freshwater wetlands (Age, geography, and physical characteristics are all variable; percentage removals are based on concentration.) [from Kadlec 1980, except for the last 3 entries]

	Loading People/ Acre		Nitrogen				Phos-	Sus-	Coliforms	
			BOD	NO ₃ -	NH ₄ +	TN	phorus Total	pended - Solids	Fecal	Total
Houghton Lake, Michigan	3	1		40	99		97**			
Bellaire, Michigan	11	. 2				91*	97*	*		
Mt. View, California	1600	4	<u>_19</u>	54	39		3	— 250		
Hay River, N.W.T.	40	5	98	••	96	• 5 • • • 5	98	97		99
Wildwood, Florida	5	20		99	99	90	98	٠	99	
Kincheloe, Michigan	15	20		99	1.98		••			
Brillion, Wisconsin	7	55	80	51	••		13	29	••	86
Dundas, Ontario		59	80	83	96	88	87,98*	*47	66	14
Great Meadows, Massachusetts	* * _ :-	69	89	33	98		. 89	12	66	—1285
Gainesville, Florida Ewel and Odum (1978) A I.	* * *s					881			
Clermont, Florida Zoltek et al. (1979)		•	**	98,.5	88	43*	**			
Oshkash, Wisconsin Spangler et al. (1978)	1) 10, 11		91—97			•	75 °	•••		Tale 1

^{*} Total dissolved nitrogen

Kadlec 1980, Zoltek et al. 1979, and Ewel and Odum 1978. The data clearly show that some areas have been able to process wastewater for long periods. It should be noted, however, that in only one instance (Mt. View, California), was the loading rate very high and the efficiency of that system for removal of BOD, NO₃-, NH₄+, P and suspended solids was very low or nonexistent. It is not known whether the loading rate is simply too large or whether the wetland capacity has already been exceeded.

Other long-term issues must be considered. Tilton and Kadlec (1979) suggested that long-term application of wastewater to peat wetlands would likely result in the replacement of bog species by other plants, especially cattail, that are more common in nutrient rich environments. There are, however, few data to support this hypothesis. Whigham et al. (1980) found that the freshwater tidal wetlands that they were studying, underwent dramatic species shifts during the first year of application of chlorinated secondarily treated wastewater. Those effects might be alleviated by varying the contact time, but it is also interesting to note how vegetation can respond to stress situations.

Figure 1 shows seasonal trends of biomass, nitrogen (Total N), and phosphorus (Total P) for one site for two years of their study. Site 2 received wastewater for two daily

^{**} Total dissolved phosphorus

^{***} Includes organic matter

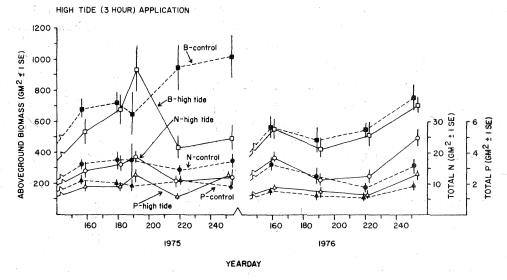


Fig. 1. Seasonal patterns of standing crop biomass, Total N, and Total P for Site 2 which received 12.5 cm of wastewater daily during two 3-hour spray periods (Whigham et al. 1980). Data for control Site 8 are shown for comparison. All values are means \pm 1 standard error of the mean.

periods of three hours of application at a rate of approximately 12.5 cm per day. There was a sharp decline in all three variables by the fourth sampling date. This response was due to elimination of almost all annuals (Whigham et al. 1980) which are very highly productive (Whigham et al. 1978). Biomass and Total N remained low throughout the first growing season. In 1976, the dramatic early season decline in the three variables did not occur. Although the annuals were still absent or uncommon, biomass and standing stocks of N and P remained high because the perennials did not experience the mid-summer die-back that normally occurs in those wetlands (Whigham el al. 1978). There were distinct species shifts during their study, yet the entire system responded to ameliorate those shifts. This leads to the question of whether or not shifts in species composition affect any important functional processes.

Zoltek et al. (1979) also found short-term changes in vegetation in the Florida wetland that they studied. Shrub species, Hibiscus, had very high growth rates during a dry year while floating species (e.g., Lemna) were abundant during wet years. They suggested that the changes may, in fact, be beneficial because it permits the wetland to respond to variations in hydrologic as well as nutrient loading rates. Odum and Ewel (1978) found floristic changes in Cypress stands following wastewater addition, but the long-term effects are unknown.

Few data exist on the fate of heavy metals, toxic chemicals, and pathogenic organisms in freshwater wetlands (Kadlec 1980), although (Valiela et al. (1974, 1976), and Banus et al. (1935) have performed extensive experimentation in salt marshes and have found very few negative effects as a result of long-term addition of sludge.

What kind of changes occur in wetland substrates after long-term application of wastewater? Odum and Ewel (1978) found significant differences in substrate composition in Florida Cypress domes while other wetlands (Kadlec 1980) appear to be able to assimilate wastewater for long periods of time without any significant changes in the substrate. Kadlec reported, however, one instance where the substrate have changed and

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resembled sewage sludge. Whigham and Simpson (unpublished data) found similar results in the freshwater tidal wetland that they studied. Zoltek et al. (1979) have suggested that substrate, as well as other ecosystem components, changes can be minimized by using rest periods between wastewater additions.

COST-BENEFIT ANALYSIS

The most detailed cost-benefit analyses have been performed by Odum and his his colleagues at the University of Florida (Ewel and Odum 1978). As an example, they calculated that 1,000 gallons of wastewater can be treated for \$0.42 compared to \$1.07 for advanced wastewater treatment and \$0.63 for upland spray irrigation (1978 dollars). Their work has resulted in consideration being given to widespread use of Florida wetlands in situations where the amount of wastewater to be treated is not excessive. Similar planning is being considered in Michigan where peat-based wetlands may be used to treat wastewater from small municipalities. It does not appear that wetlands can be used to process wastewater from large cities, although wastewater polishing may be possible (Whigham et al. 1980). It may also be possible to use wetlands to treat non-human wastes. Turner et al. (1976) have, for example, shown that it would be economically feasible to use wetlands to polish wastes from menhaden processing plants in Louisiana.

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