

Research article - CORRECTED PROOF**A new coral species introduced into the Atlantic Ocean - *Tubastraea micranthus* (Ehrenberg 1834) (Cnidaria, Anthozoa, Scleractinia): An invasive threat?**Paul W. Sammarco^{1*}, Scott A. Porter² and Stephen D. Cairns³¹Louisiana Universities Marine Consortium (LUMCON), 8124 Hwy. 56, Chauvin, LA 70344 USA²EcoLogic Environmental Consulting, 1542 Barrow St. Houma, LA 70361 USA³Department of Invertebrate Zoology, Smithsonian Institution, P. O. Box 37012, NMNH, W-326, MRC-163, Washington, D. C. 20013-7012, USAE-mail: psammarco@lumcon.edu (PWS), Ecologic2020@aol.com (SAP), cairns@si.edu (SDC)

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Abstract

Over the past 60-70 years, the invasive Indo-Pacific coral *Tubastraea coccinea* (Lesson 1830; Cnidaria, Scleractinia) has colonized the western tropical Atlantic Ocean - the Americas, the Antilles, northern Gulf of Mexico (GOM), and many of its 3,600 oil/gas platforms. It is now the single, most abundant coral on artificial substrata in the GOM, with hundreds of thousands of colonies on a single platform. Here, we report for the first time the observation of a closely related congener in the western Atlantic - the Indo-Pacific azooxanthellate *Tubastraea micranthus* (Ehrenberg 1834) - and suggest that it may pose a threat similar to *T. coccinea*. A total of 83 platforms, including deep-water, toppled, Rigs-to-Reefs structures, were surveyed in the northern Gulf of Mexico between 2000 and 2009, from Matagorda Island, Texas to Mobile, Alabama, USA, between the depths of 7 and 37 m, by SCUBA divers. Five platforms were surveyed by Remotely Operated Vehicle (ROV) to depths of up to 117 m. *T. micranthus* was found on only one platform - Grand Isle 93 (GI-93), off Port Fourchon, Louisiana, near the Mississippi River mouth, at the cross-roads of two major safety fairways/shipping lanes transited by large international commercial ships. The introduction appears to be recent, probably derived from the ballast water or hull of a vessel from the Indo-Pacific. If the growth and reproductive rates of *T. micranthus*, both sexual and asexual, are similar to those of *T. coccinea*, this species could dominate this region like its congener. It is not known whether this species is an opportunist/pioneer species, like *T. coccinea*, a trait protecting benthic communities from its dominance. The question of rapid-response eradication is raised.

Key words: coral, *Tubastraea coccinea*, *Tubastraea micranthus*, introduction, invasive species, Atlantic Ocean, Indo-Pacific**Introduction**

In recent times, intentional species introductions have been of concern because of their secondary impacts on the natural environment and local species (Secord 2003). Of even greater concern are accidental introductions that have proven to be detrimental to ecosystems that they have colonized (Roberts and Pullin 2008). Invasive marine species are of particular concern because of the ease with which they can disperse and colonize nearby habitats once they have established a new population (Griffiths 1991; Johnson and Carlton 1996; Wonham et al. 2000). Vectors of transport for invasive marine or freshwater species (Kerr et al. 2005) include ballast water of barges or ships (Chesapeake Bay Commission 1995; ICES 2002), the hulls of the same (Minchin and Gollasch 2003), the towing

of oil and gas platforms to new sites (Hicks and Tunnell 1993), accidental release of exotic species from mariculture operations (Sapota 2004), and deliberate release of exotics by aquarium hobbyists (Weidema 2000; Christmas et al. 2001; Hindar et al. 2006).

An example of an introduced species widely distributed from the northern temperate waters through western Atlantic tropical coral reefs (Chapman 1999) is the Japanese/Siberian marine alga *Codium fragile* (Suringar; Hariot 1889), now common throughout the western Atlantic (Trowbridge 1998; Pederson 2000; Williams 2007). Another example is the Indo-Pacific volitan lionfish (*Pterois volitans* Linnaeus 1758), released into western Atlantic waters ≥ 10 years ago (Whitfield et al. 2002; Hamner et al. 2007) and now distributed from New York through the Caribbean.

Recently, there has been concern about invasive species occurring in the Gulf of Mexico (Osman and Shirley 2007). Species of interest include the Australian scyphozoan *Phyllorhiza punctata* (von Lendenfeld 1884), which have occurred in high enough densities (Perry and Graham 2000) to dramatically suppress ichthyoplankton and invertebrate larval densities (Graham et al. 2003; Graham and Bayha 2008). *Didemnum perlucidum* (Monniot 1983), an Indo-Pacific colonial encrusting didemnid ascidean, was also introduced to the Gulf during the 1990s and has proven to be an impressively strong competitor for space with other sessile epibenthic organisms in this region (Culbertson and Harper 2002; Lambert, 2002; Sammarco 2007).

There are very few reports of successful introductions of corals into the Atlantic. The most striking example and that of greatest concern has been the Indo-Pacific species *Tubastraea coccinea* Lesson, 1829. It was first recorded in 1943 in Puerto Rico and then in 1948 in Curacao, Netherlands Antilles, where they were collected on ship's hulls (Cairns 2000). By the late 1990s and early 2000s, the species was reported to occur off Belize and Mexico (Fenner 1999); Venezuela, northern Gulf of Mexico, and the Florida Keys (Fenner 1999; Fenner and Banks 2004; Sammarco et al. 2004); Brazil (Figueira and Creed 2004); Colombia, Panama, the Bahamas, and throughout the Lesser and Greater Antilles (Humann and DeLoach 2002). *Tubastraea coccinea* is now a highly successful invasive species in the western Atlantic, being the single most abundant scleractinian coral, hermatypic or ahermatypic, in the northern Gulf of Mexico on artificial substrata – by orders of magnitude (Sammarco et al. 2004, 2007). Hundreds of thousands of colonies occur on individual oil and gas platforms there, in densities up to 36 m⁻² (Sammarco 2008). It does occur on deep banks in the northern Gulf of Mexico but has not been observed to be abundant there (Schmahl 2003; Hickerson et al. 2006; Schmahl and Hickerson 2006). No studies have yet been conducted to investigate whether or not *T. coccinea* has expanded its populations at the expense of other species. The lack of such an active competitive exclusion in the wake of such a dramatic population explosion would presume that the ecosystem had one or more open niches. This is also not known.

A second incident regarding the accidental introduction of a coral to the Atlantic is that of *Fungia scutaria* (Lamarck; Vaughan 1907) to Discovery Bay, Jamaica, W.I., observed as early

as 1973 (P.W. Sammarco pers. obs; J. Lang pers. comm.; LaJeunesse et al. 2005). In this case, it is likely that this species will remain “non-invasive”. This is because it is a vagile, epibenthic coral and has probably invaded the open niche of soft bottom in that environment, where there is no previously existing coral species with similar habitat requirements.

Together and individually, we (Sammarco and Porter) conducted surveys on the distribution and abundance of scleractinian corals on numerous platforms throughout the northern Gulf of Mexico – from the waters off Corpus Christi, Texas to those off Mobile, Alabama. We encountered numerous coral species during these surveys (see Sammarco et al. 2004; Sammarco 2008); additional results from those surveys will be published elsewhere. Here, however, we report for the first time the introduction of a new species to the Gulf of Mexico, and to the Atlantic Ocean – the Indo-Pacific species *Tubastraea micranthus* (Ehrenberg 1834) (Cnidaria, Dendrophylliidae, Cairns, 2001). *T. micranthus* is a close congener of *T. coccinea*, occurring in shallow water. *T. micranthus* is considered to be unique among the 1500 scleractinian species because, although it is azooxanthellate, it is a hermatypic coral, being capable of building reef-like structures in the Pacific (Roberts et al. 2009). It is also known to have exhibited population expansion characteristics in new environments elsewhere in the world (Loch et al. 2004).

Here we define “invasive species” as “an alien species whose introduction does or is likely to cause economic or environmental harm or harm to human health” (US Exec. Ord. 13112, 1999), or alternately “a species that is non-native to the ecosystem under consideration and whose introduction causes or is likely to cause economic or environmental harm or harm to human health” (Invasive Species Advisory Comm. 2005). Such species often spread, encroach upon, or take over the habitat of native species. An “alien species”, on the other hand, is “any species, including its seeds, eggs, spores, or other biological material capable of propagating that species, that is not native to that ecosystem” (ibid.). Not all introduced or alien species are necessarily invasive. Carlton and Eldredge (2009) present evidence that out of the > 400 introduced or cryptogenic species now reported to exist in Hawaii, < 10% may be considered invasive. Our studies show that *Tubastraea micranthus* is an introduced species that has the potential of becoming an invasive species. This

is because of its likelihood of exhibiting population expansion characteristics similar to its close Indo-Pacific congener, *Tubastraea coccinea*, which has experienced a remarkable spread throughout the western Atlantic in recent decades.

We report the results of surveys in both shallow and deep water, via SCUBA and ROV reconnaissance, regarding the occurrence of *Tubastraea micranthus* on oil and gas platforms. We also consider the ecological implications of this recent species introduction to the Gulf of Mexico.

Materials and methods

Study sites

Eighty-two (82) oil and gas platforms were surveyed for the presence of scleractinian corals in four sectors across the continental shelf of the northern Gulf of Mexico, from Corpus Christi, TX to Mobile, AL (Annex 1; Cairns and Zibrowius 1997; Sammarco 2005, 2006, 2007, 2008, in prep.; Sammarco et al. 2006, 2008, 2009). Most were standing production platforms; some were toppled platforms within the federal/state “Rigs-to-Reefs” artificial reef program (see Dauterive 2000). One series of surveys covered the offshore lease areas entitled High Island (HI), West Cameron (WC), and East Cameron (EC) to a depth of 10-30 m via SCUBA between 2000 and 2003. A second broader series of four (4) cross-continental shelf transects were performed in the Matagorda Island (MI), Brazos (BA), High Island (HI), West Cameron (WC), Garden Banks (GB), South Timbalier (ST), Ship Shoals (SS), and Main Pass (MP) sectors. These were done between 2003 and 2007, covering 10-37 m depth. A third survey was performed within the Grand Isle (GI) and South Timbalier (ST) lease areas to a depth of 30m (Porter) from 2004-2009. A final survey was done in the South Timbalier (ST) and East Cameron (EC) sectors in depths of 33-117 m. Additional deep surveys were performed on “Rigs-to-Reefs” structures via Remotely Operated Vehicles (ROVs) in the High Island (HI), West Cameron (WC), and East Cameron (EC) lease areas in 2007-2008. (See Annex 1 for a listing of all sites and their locations.) Most sites were surveyed only once.

Survey techniques

A total of 41 SCUBA divers spent ~1,120 person-hours underwater surveying the submerged jackets of these platforms. Techniques of

recording and assessing coral species distribution, abundance, density, diversity, etc. are reported in detail in Sammarco et al. (2004); the reader is referred there for further information. In summary, personal visual surveys were implemented using visual belt transects. The transects covered the width of the cylindrical steel platform supports (horizontal, vertical, and diagonal), and, a log₅ scale was used to estimate coral abundance. The area surveyed was calculated from architectural schematics of the platforms provided by the oil and gas companies and used to estimate coral densities.

Continuous belt-transects were also recorded on digital video tape via a diver-held underwater digital video camera. In addition, we used two Remotely Operated Vehicles (ROVs) to conduct our deeper surveys – a SeaBotix LBV-300 and a Phantom 2. Each was fitted with vertical and horizontal propulsion units, site-to-surface video units with a topside monitor, lights, and a sample-retrieval unit (manipulator arm) for collection purposes. Representative coral specimens were collected, returned to the laboratory, cleaned, sun-bleached and dried for purposes of a more detailed examination and verification of species identification. Live specimens of the target species were also collected by Porter.

Several other deeper, toppled platforms that had been decommissioned, extricated, placed on a barge, and transported to shore for salvage, were also examined onshore (Porter); (the sites of origin of these platforms were uncertain).

Video images were captured on mini-DV or DVD via several types of units. A SONY FX-7 HD camera was used in conjunction with an HP 6130 laptop computer possessing a duo-core processor and Pinnacle-12 HD software. A high-resolution 3-chip color SONY camera was also used to record images received from the ROV, in conjunction with a Dell XPS laptop fitted with a duo-core processor using MicroSoft video imaging software.

Processing of visual information on the DVDs was performed with the assistance of a Dell Precision 340 desktop computer with a Pentium 4 processor, fitted with several pieces of image analysis software - Nero 7.0, VideoLAN, and MicroSoft Windows Media Player, capable of zoom and still-image capture. Data were logged in EXCEL files.

Species identification was verified by one of us (Cairns) using reference specimens held at the Museum of Natural History, Smithsonian Institution, Washington, D.C.

Results

Of the 82 structures surveyed, only one was found to possess *Tubastraea micranthus* - the standing production platform GI-93-B (28°32.96'N, 90°40.11'W; Figure 1). The first observation was made in 2006 (Porter). The first way in which these corals were distinguished from *Tubastraea coccinea* was by color. The colonies were dark green or black in tissue pigmentation, as opposed to the bright orange-yellow pigmentation of *T. coccinea* (the origin of its common name “orange cup coral”, “sun polyp”, or “sun coral”; Humann and DeLoach 2002; Figure 2). The tentacles of *T. micranthus* were sometimes grey in color. (The pigmentation is derived from the tissue, because this species, like its congeners, is azoo-xanthellate.)

The second distinguishing character was its colony morphology. The stalk of the colony sometimes extended vertically > 0.5 m above the substratum. In addition, the corallites were often 20 mm longer than those of *T. coccinea* (Figures 2 and 3). Some colonies possessed 30 cm-long branches, with a pattern of alternating corallites (see Cairns and Zibrowius 1997 for a complete taxonomic description of this species). Calices of *T. micranthus* are smaller than those of *T. coccinea*, the former being 6-8 mm in diameter, and the latter 10-11 mm. The calical structure also varied slightly between the two species, with the tertiary septa of *T. coccinea* sometimes being well developed and having lacinate axial edges (Figure 4).

The colonies were found *in situ* on GI-93-B at depths ranging from 18-22m (Figure 1). In all cases, the colonies were alive. In some cases, colonies were returned to shore and kept alive in aquaria for up to 3 years and, at the time of writing, are still being held successfully in captivity. Some variation in colony morphology was observed, although it was not beyond that described to occur in Indo-Pacific specimens (Sammarco et al. 2008). An examination of dried and bleached specimens of both *Tubastraea coccinea* and *T. micranthus* confirmed morphological differences between the skeletons of each species.

High abundances of this species were found on GI-93-B. Initially, a small patch of scattered colonies was observed in 2006 to extend ~3 m in length along a horizontal support. In 2009, this same patch was observed to have extended over 30 m on the same support, with ~80% live cover in some areas. *Tubastraea micranthus* was clearly expanding its cover in this benthic

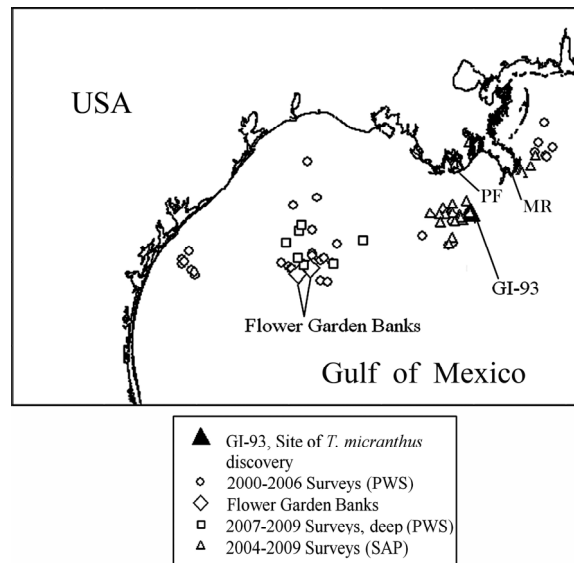


Figure 1. Map of the northern Gulf of Mexico, depicting survey sites for scleractinian corals (see Annex 1 for list of survey sites and geographic coordinates). Surveys conducted in several sets – from 2000 to 2006 across the shelf by SCUBA by Sammarco’s research team, from 2004 to 2009 in the central and near-eastern sector by Porter’s team, and deep-water surveys from 2007-2009 in the near-western and central sectors by SCUBA and ROV by Sammarco’s team. *Tubastraea micranthus* found only on GI-93. MR – Mississippi River mouth, PF - Port Fourchon.

community, apparently at the expense of other sessile epibenthic organisms, but not necessarily to their total exclusion. No quantitative data are yet available upon which to draw conclusions on this species’ differential interspecific competitive capabilities. No massive coverage similar to that of *T. coccinea* has yet been observed on platforms in this region.

Discussion

The fact that *Tubastraea micranthus* was observed at only one site within the entire survey area covering the northern Gulf of Mexico suggests that this is a very recent introduction. Also, this species is known to grow to a height of 1.0 m. The maximum colony height observed here was 0.5 m, indicating that these colonies might be relatively young. It is known from parallel studies that *T. coccinea* is a species that is capable of successfully dispersing its larvae over large distances and recruiting there (Sammarco 2005, 2006, 2007, 2008, 2009, in prep.; Fenner 1999, 2001; Fenner and Banks 2004; Figueira and Creed 2004; Humann and DeLoach 2002; Sammarco et al. 2006, 2009).

Figure 2. (A) *Tubastraea micranthus* colonies, identified by their dark tissue coloration and extended vertical growth, shown at 20m depth on a horizontal support of the oil platform GI-93 in the northern Gulf of Mexico. The coral species surrounding these colonies are *T. coccinea* (photo by SAP); (B) *Tubastraea coccinea* colonies observed on an oil platform in the High Island (HI) lease area. Note the more compressed habit and lighter, brighter tissue color (photo by PWS).

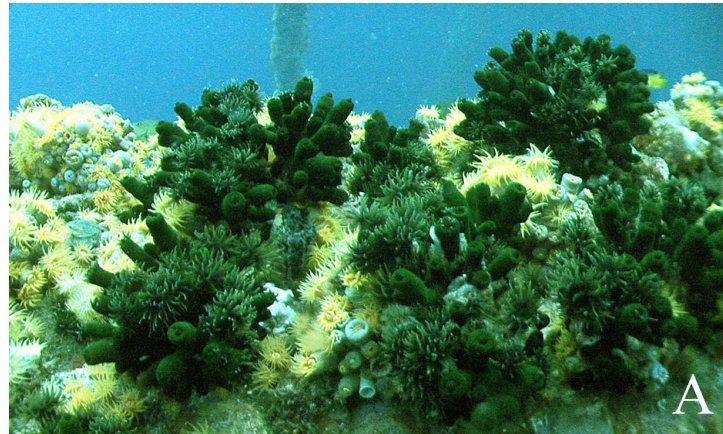


Figure 3. (A) Colony of *Tubastraea micranthus* that has been cleaned of its tissue, dried, and bleached. Note the extended corallites and distinct sparse branching pattern. (B) Colony of *T. coccinea*, similarly prepared. Note the compressed habit and density of short corallites (photos by SAP).

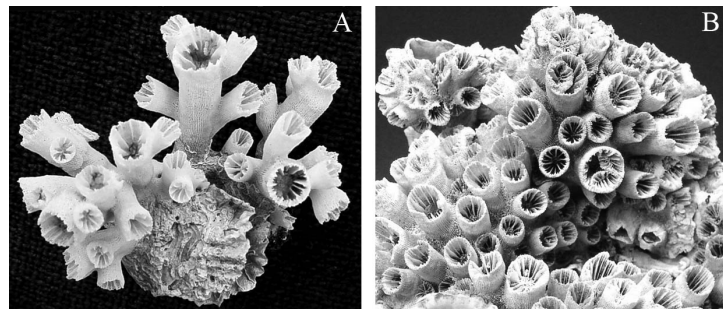
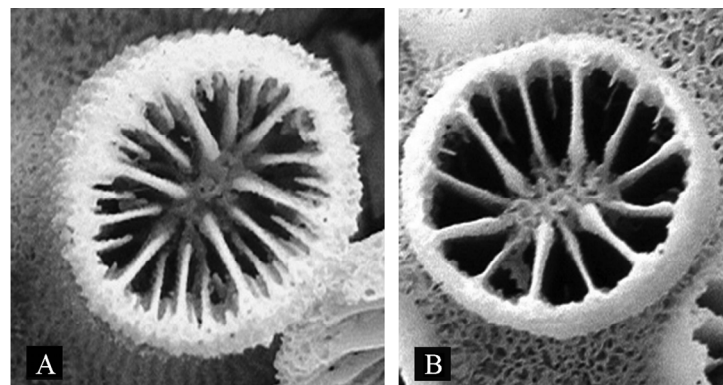


Figure 4. (A) Single corallite of *Tubastraea micranthus*, showing details of the calice structure. Note extensive development of the secondary and tertiary septa. (B) Single corallite of *Tubastraea coccinea*, showing comparative details. Note extensive development of the secondary septa, but only nominal development of the tertiary ones (photos by SAP).



We believe that, because the site at which *T. micranthus* was observed is in relatively close proximity to the mouth of the Mississippi River (near New Orleans, Louisiana) and Port Fourchon, a major deep-water port for the northern Gulf of Mexico, it is quite possible that the vector for this introduction was a vessel traveling from the Indo-Pacific.

If this species has been observed on this (or other) platform(s) before, it is possible that it may have been mistaken for a color morph of *Tubastraea coccinea* by SCUBA divers and other underwater surveyors and collectors, whether observing personally or *via* ROV. Until now, apparently no specimens of *T. micranthus* have been reported or submitted to taxonomic experts for formal identification from this region.

Since *Tubastraea coccinea* was introduced into the western Atlantic in 1943, 56 yrs from the time of writing, it has become the single, most abundant scleractinian coral in the northern GOM. This raises concern over the introduction of its congener – *T. micranthus*. It is possible that *Tubastraea micranthus* may represent the same threat of spread in the GOM and the western Atlantic as *T. coccinea*. It will be critical to monitor its population dynamics in this region to determine whether it has the same capability to expand its population as its congener. *T. coccinea* is known to reproduce asexually via asexual planula production (Ayre and Resing 1986; Shearer 2008) and the highly effective production of runners to produce new colonies (Vermeij 2005; Pagad 2007). It also produces planulae sexually (Glynn et al. 2008a, b). If the reproductive capacity and larval dispersal capabilities of *T. micranthus* are similar to those of *T. coccinea*, *T. micranthus* could potentially reach similar high abundances in the western Atlantic with time (Sammarco et al. 2004; Shearer 2008). Indeed, we predict that if its reproductive, dispersal, and colonization capabilities are similar, there could be a massive biogeographic expansion of this species throughout the Gulf and into the western Atlantic over the next 20-40 years. Our preliminary observations suggest that *T. micranthus* may be competitively dominant over its predecessor, *T. coccinea*, and *T. coccinea* is already known to be a competitive dominant in its benthic community. This implies that *T. micranthus* could represent a threat to other sessile, epibenthic marine fauna and flora in western Atlantic tropical and subtropical waters.

The appearance of *Tubastraea micranthus* represents the third documented case of a scleractinian coral which has been introduced to the western Atlantic. The first was *Fungia scutaria*, introduced to Discovery Bay, Jamaica during the late 1960s (Sammarco pers. obs; J.C. Lang, pers. obs; see LaJeunesse et al. 2005). Specimens of this Indo-Pacific species were apparently transported to Jamaica and kept in seawater aquaria at the Discovery Bay Marine Laboratory. These, however, were open-system running seawater aquaria, and most likely allowed planula larvae to be carried out to the lagoon through the effluent discharge pipes. Adults may also have been introduced into the lagoon directly. The second successful introduction of a scleractinian coral was that of *Tubastraea coccinea*. That introduction has been described in detail by Fenner (2001), Cairns (2000), and Fenner and Banks (2004; see above).

The question arises as to whether a rapid-response is called for on the part of government to eradicate this potential invasive in its early stages of colonization, rather than waiting too long. In the early stages of colonization, successful complete eradication of an invasive species is possible, as has been successfully effected with the sabellid polychaete *Terebrasabella heterouncinata* (Fitzhugh and Rouse 1999). This species was an ectocommensal of South African abalone, which was imported to a central coastal town in California for mariculture purposes (Culver and Kuris 2000). Eradication was effected by removing all of their preferred hosts (a turban snail) and screening the release of potentially contaminated materials from the mariculture unit. A similar rapid response to the infestation of *Caulerpa taxifolia* shortly after its introduction to southern California has also resulted in the near-elimination of this species (Anderson 2005; Woodfield and Merkel 2006). Delay of eradication of invasive species, however, is much more difficult to achieve (Simberloff 2000; Hewitt et al. 2005) and potentially problematic. A community may have reached a new stable equilibrium, integrating the new species. Under such conditions, if the species is removed, the community can become destabilized and rapidly thrown into disequilibrium. This has been documented to occur in the terrestrial environment (Bergstrom et al. 2009; Casey 2009) and could potentially occur in marine or freshwater environments.

Additional surveys within the Gulf of Mexico, including greater depths, should be performed. In addition, investigators conducting similar surveys in other parts of the Atlantic are encouraged to familiarize themselves with *Tubastraea micranthus*, its morphological features, and how they vary from those of *Tubastraea coccinea*, and be aware that this species might appear in their study sites. If *Tubastraea micranthus* is to be eradicated from the Gulf of Mexico, such action should be taken soon.

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Annex 1. List of the 82 oil and gas platforms surveyed for scleractinian corals between 2000 and 2009. Most were standing production platforms. Some were toppled platforms within the “Rigs-to-Reefs” Program. Most were surveyed by SCUBA, between 7 and 37 m depth. Several were surveyed by ROV to a maximum depth of 117 m. Platforms divided into transect survey sectors within the northern Gulf of Mexico, covering the western, near-western, central, and near-eastern regions. Latitudes and longitudes of survey sites also shown. PWS = Sammarco; SAP = Porter.

Platform	lat.	long.	Platform	lat.	long.	Platform	lat.	long.
Far Western Sector (2003-2006, PWS)			Central Sector (2003-2006, PWS)			Near-Eastern Sector (2003-2006, PWS)		
BA-104-A	27.8669	-96.0334	SS-277-A	28.2993	-91.0876	MP-144-A	29.2924	-88.6691
BA-A-133-A	27.8545	-96.0364	ST-163-A	28.5720	-90.4996	MP-159-1	29.6491	-88.4647
BA-133-D	27.8388	-96.0282	ST-165-A	28.5767	-90.5769	MP-236-B	29.4054	-88.5844
MI-A-4A	27.9042	-96.0892	ST-188-CA	28.5010	-90.3808	MP-265-A	29.3467	-88.2816
MI-618-A	28.0222	-96.3071	ST-190-A	28.4663	-90.4461	MP-288-A	29.2398	-88.4095
MI-651-A	28.0222	-96.3071	ST-292-A	28.2141	-90.4203	MP-289-B	29.2585	-88.4415
MI-672-A	27.9942	-96.2596	ST-295-A	28.1963	-90.5413			
MI-672-B	27.9688	-96.2909						
Western Sector (2000-2003, PWS)			(2004-2009, SAP)			(2004-2009, SAP)		
EC-317B	28.2084	-92.9519	GI-90	28.5850	-90.2383	MC-387	28.5583	-90.9317
HI-A-30A	28.0962	-93.4784	GI-93**	28.5583	-90.0800	MP 296	29.2483	-88.6650
HI-A-349B	28.0703	-93.4691	GI-94	28.5400	-90.2750	MP 306	29.195	-88.5550
HI-A-368B	27.9620	-93.6709	GI-95	28.5250	-90.1317	MP-313-B	29.115	-88.7683
HI-A-370A	27.9855	-93.4584	GI-106	29.3983	-90.0967	SP-62	29.0583	-88.9650
HI-A-376A	27.9620	-93.6709	ST-131	28.7083	-90.1467	SP-72	29.0433	-88.9450
HI-A-382	27.9132	-93.9351	ST-138	28.6300	-90.4100			
HI-A-385C	27.9168	-93.9168	ST-139	28.6667	-90.4550			
HI-A-386B	27.9717	-93.5178	ST-146	28.6233	-90.6483			
HI-A-568B	27.9763	-94.1439	ST-150	28.6250	-90.2567			
HI-A-571	27.9556	-94.0271	ST-156	28.5750	-90.1950			
WC-618	28.0367	-93.2312	ST-162	28.5567	-90.4450			
WC-630A	28.0034	-93.2941	ST-171	28.5483	-90.6217			
WC-643A	27.9814	-93.0342	ST-172	28.5300	-90.5750			
			ST-178	28.5433	-90.2883			
(2003-2006, PWS)			ST-179	28.5183	-90.2967			
GB-189-A	27.7786	-93.3095	ST-185	28.4817	-90.2367			
GB-236-A	27.7611	-93.1377	ST-186	28.5017	-90.2817			
HI-A-237-A	28.6678	-93.8857	ST-195	28.5300	-90.6850			
HI-A-287-A	28.3610	-93.7690	ST-196	28.5417	-90.7317			
WC-312-1	29.1872	-93.5822	ST-198	28.4553	-90.7050			
WC-414-A	28.7579	-93.3878	ST-204	28.4617	-90.3983			
WC-522-A	28.3759	-93.4912	ST-265	28.2733	-90.4633			
			ST-295	28.1963	-90.5413			
(2007-2008, PWS)								
(Shallow - < 33m - and deep ROV surveys)								
HI-A-376	27.9620	93.6709						
WC-643-A	27.9814	93.0342						
EC-273AP R2R*	28.2550	-92.3930						
HI-A-271 R2R*	28.4395	-93.7169						
HI-A-281 R2R*	28.3642	-93.7853						
HI-A-355R2R*	28.3642	-93.7853						
HI-A-492 R2R*	28.2266	94.0583						

*Toppled Rigs-to-Reefs platforms; all others are standing production platforms.

**Site of observation of *Tubastraea micranthus*.